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The rate of lindane biodegradation in different constructed wetlands

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ABSTRACT

In the last decades, the use of some types of pesticides was restricted or completely banned because of their adverse secondary effects to the environment and human health. The harmful effect of an insecticide y-hexachlorocyclohexane (y-HCH, lindane) on the environment and organisms is well known, although field application was banned in 2010, and contamination of soil and water is alarming. The aim of this two-year study was to compare removal of γ-HCH (Lindane) and its isomers in different constructed wetlands at a heavily contaminated site called Hajek.

RESULTS AND DISCUSSION

The genes confirming the presence of enzymes involved in aerobic lindane degradation were detected in almost all of the samples (Table 1). The



INTRODUCTION

The gamma isomer of 1,2,3,4,5,6-hexachlorocyclohexane (y-HCH), also known as lindane, is a cyclic, saturated and highly chlorinated pesticide. One of the serious environmental burdens in Czech Republic is connected with former production of lindane and technical HCH in the body of the dump of mine near the village Hájek (south-eastward from Ostrov in region of Karlovy Vary). Nowadays there are several constructed wetlands on site to minimise lindane contamination of environment.

MATERIALS AND METHODS

abundance of genes involved in lindane biodegradation was similar in the wetlands A and B, but rather low in the wetland C.

The NGS results indicated dynamic changes in bacterial community in 2017 and 2018, probably due to very dry season in the latter year. Sulfur and iron reducing bacteria dominated in the inflow and outflow of wetlands in 2017 (Fig. 2). In 2018, sediment from the B wetland, samples Desulfovibrio dominated. This species is known to be capable of lindane dechlorination. Microbial community in the outflow from B wetland in 2018 was more diverse comparing to previous year results (Fig. 3).

Interestingly, in C output mostly Gallionella dominated and Sulfuricurvum species, while in the sediment presence of this genera is minimal. But sediment samples turned out to be rich in Acinetobacter, Comamonas, Flavobacterium, and Parapusillimonas. Generally microbial diversity in inflow samples significantly higher comparing to outflow samples from all wetlands regardless of the year.

Fig. 2: Relative representation of microorganisms at family level (abundance > 5%) in the wetland B inflow and outflow and sediment samples in 2017.



Fig. 3: Relative representation of microorganisms at family level (abundance > 5 %) in the wetland B inflow and outflow and sediment samples in 2018.

Wetland A mimicked natural shallow-water wetland, wetland B was filled with organic substrate to support biodegradation, and on the inflow of wetland C, there were three permeable reactive barriers filled with iron sponges, wetland E filled with biochar (Fig.1). Specific genes were assessed for their potential in lindane biodegradation in all wetland biotops. The water and sediment samples were collected for DNA extraction and analysis by real-time quantitative PCR (qPCR) to determine total bacterial biomass, and presence of functional genes (LinA, LinB, LinB-RT, LinD). Based on the qPCR results, water and sediment samples from one part of the wetland system (A,B, C and E) were chosen to be analyzed by nextgeneration sequencing (NGS) to estimate the bacterial diversity.

Chemical analysis data (Fig 4) of wetland in- and out-flow indicate partial HCH elimination from the drain water. Wetland C is most effective in HCH removal.



Fig. 4: The sum of HCH concentration in inflow and A,B,C outflow samples in 2018.

Different bacterial diversity was found in sediments and water in all wetlands. Obviously, aerobic conditions in water were much more favorable for lindane-degrading bacteria in comparison with the more anaerobic conditions in sediments.

Bacterial groups and enzymes involved in biogeochemical cycles of sulphur and iron were identified in large quantities. According to chemical data we can conclude that wetland C and B have higher efficiency to lindane removal comparing to natural wetland, while in A and B wetlands the biodegradation markers were more abundant.





Fig. 1: Scheme of constructed wetland at the locality Hajek. Wetland A mimicked natural shallow-water wetland, wetland B was filled with organic substrate to support biodegradation, and on the inflow of wetland C iron sponges were installed.

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